## Western Alaska Salmon Stock Identification Program

## Technical Document: ${ }^{1}$

Title: Status of the SNP baseline for sockeye salmon
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## Introduction

The single nucleotide polymorphism (SNP) baseline for sockeye salmon that will be used for mixed stock analysis (MSA) to estimate stock contributions of catches sampled under the Western Alaska Salmon Stock Identification Program (WASSIP) is in a state of perpetual improvement. The collections that make up this baseline were collected over the past twenty years and were funded by many sources including the State of Alaska through general funds and disaster funds, the North Pacific Research Board, National Park Service, Federal Office of Subsistence Management, Pacific Salmon Commission, and the Exxon Valdez Oil Spill Trustee Council.

The suite of SNP markers screened for the baseline has also changed through time and will continue to grow or change as more markers become available. We currently screen for 42 nuclear and three mitochondrial markers, but the WASSIP Advisory Panel has requested that 96 SNP markers be incorporated into the baseline to improve the precision and accuracy of stock composition estimates. To meet this request, we are contracting the development of at least 50 SNP markers that are targeted to differentiate among sockeye salmon populations spawning within western Alaska and the Alaska Peninsula drainages (Technical Document 6). These new SNP markers will be assessed after screening a fraction of the baseline and the best-performing SNP markers will be added to the baseline during the winter of 2009/2010.

[^0]Here we present the current state of the baseline based on samples collected through the 2008 collection season and genotyped for the currently available 42 nuclear and three mitochondrial SNP markers.

## Methods

## Tissue Sampling

Baseline samples for SNP analyses were collected from spawning populations or obtained from existing agency archives from throughout the range of sockeye salmon in the Pacific Rim (Table 1). We used published genetic structure information (Beacham et al. 2006) to determine appropriate areas to sample outside the Bering Sea drainages. Target sample size for baseline collections was 95 individuals across all years for each population to achieve acceptable precision for the allele frequency estimates (Allendorf and Phelps 1981; Waples 1990a) and to accommodate our genotyping platform.

## Laboratory Analysis

## Assaying genotypes

Genomic DNA was extracted using a DNeasy® 96 Tissue Kit by QIAGEN® (Valencia, CA). Forty-five sockeye SNP markers were assayed (Table 2), three mitochondrial DNA (mtDNA) and 42 nuclear DNA (nDNA), using $5^{\prime}$ nuclease methods described in Seeb et al. (2009). Thirtysix assays originated from Smith et al. (2005) and Elfstrom et al. (2006). Nine new markers were developed using the methods of Smith et al. (2005) or Elfstrom et al. (2006) and sequencing fifty individuals, ten individuals collected at each of five geographic locations (Russia, Bristol Bay, Kodiak Island, Southcentral Alaska, and Southeast Alaska; Habicht et al. submitted). Individuals were sequenced in both directions, and sequences were aligned and screened for SNPs using Sequencher 4.5 software (Gene Codes Corporation).

Baseline population samples were genotyped using uniplex SNP genotyping performed in 384well reaction plates and also by using the 48.48 array (Fluidigm Corporation) where 43 of the 45 markers were assayed in sets of 48 fish and One_MHC2_190 and One_STC-410 were assayed on the 384 -well platform. With either platform, genotypes from generally 384 fish were visualized using the GeneMapper (uniplex platform; Applied Biosystems) and BioMark (array platform; Fluidigm Corporation) software programs and scored for each marker by two people simultaneously. Scores were entered and archived in the Gene Conservation Laboratory Oracle database, LOKI.

## Quality control

Three measures were taken to ensure quality control of the baseline data:

1. Re-genotyping of samples - Eight percent of each collection was re-genotyped for all markers to ensure that genotypes were reproducible, to identify laboratory errors, and to measure rates of inconsistencies during repeated analyses on the uniplex and array platforms. We report here error rates for a representative baseline project which consisted of 87 baseline collections comprising 7,593 individuals ( $\sim 15 \%$ of current baseline).
2. Exclusion of individuals with an excessive rate of drop-outs - A threshold of $80 \%$ scorable markers per individual was established and all individuals that did not meet this threshold were excluded from statistical analysis and use in the baseline. This threshold was set to exclude individuals with poor quality DNA. Poor quality DNA leads to lower reproducibility and therefore adds error to the allele frequency estimates. The value of $80 \%$ was chosen based upon the observation that many individuals with high quality DNA had some dropouts, but generally less than $20 \%$ of markers, while those with poorquality DNA had higher drop-out rates. As a result, there was little difference in which individuals were excluded from analysis when picking the threshold as long as it was within the $70 \%$ to $90 \%$ range.

This rule (referred to as the " $80 \%$ rule") will also be used for samples from fishery harvests to decrease errors and estimate variances caused by poor quality DNA and missing data. This approach is an attempt to balance the benefits from better data with the loss of power to accurately and precisely estimate stock proportions due to smaller sample sizes. One other potential disadvantage of this approach is the potential to introduce another form of bias if fish that are removed from analyses are not randomly distributed in the mixture. Heterogeneity in sample removal may introduce bias in subsequent estimates of stock proportions when samples with quality genotypic data are not representative of the entire harvest being sampled. We anticipate that bias will only be a concern if significant proportions of mixtures are excluded.
3. Exclusion of duplicate individuals - Finally, we searched for suspected duplicate fish within collections by identifying pairs of individuals that had identical multi-marker genotypes at 38 or more markers. If suspected duplicates were found, the second individual in each matching pair was removed from further analyses.

## Statistical analysis

## Heterozygosity and $F_{S T}$

Genotypic data were retrieved from LOKI and were used to calculate allele frequencies. Observed heterozygosity, expected heterozygosity, and $\mathrm{F}_{\mathrm{ST}}$ (Weir and Cockerham 1984) were calculated for all markers using the program GDA (Lewis and Zaykin 2001).

## Linkage disequilibrium

All pairs of nuclear markers were tested for gametic disequilibrium within each collection using GENEPOP (version 4.0; updated version of Raymond and Rousset 1995; Rousset 2008). We defined a pair of markers to be significantly out of gametic equilibrium if tests for gametic disequilibrium were significant $(P<0.01)$ for greater than half of all collections.

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When gametic linkage was significant, we produced composite genotypes by ordering the alleles within each marker alphabetically and then stringing the alleles together by marker ordered alphanumerically. Markers that did not exhibit gametic disequilibrium with any other markers and markers that were combined were defined as loci for the remaining analyses. All mtDNA markers were combined into a single locus.

## Pooling collections into populations

Collections taken at the same location at similar calendar days in different years were pooled as suggested by Waples et al. (1990). Technical Document 2 has a more detailed investigation of temporal variation among collections taken in different years at the same site and calendar time. Samples taken at the same location, but at substantially different calendar days, and samples taken from geographically proximate locations were tested for homogeneity using a chi-square test of allele frequency distributions across all loci. Groups of collections that demonstrated homogeneity ( $P>0.01$, not corrected for multiple tests) were pooled. The pooled and the remaining unpooled collections were defined as populations in further analyses. Our protocol was to drop populations from further analyses if they were represented by sample sizes of less than 80 fish.

## Hardy-Weinberg equilibrium

Genotype distributions within collections were tested for deviation from Hardy-Weinberg expectation (H-W) using GENEPOP (version 4.0). These tests were repeated once collections were pooled into populations. For H-W, critical values $(\alpha=0.05)$ were adjusted for multiple tests within markers among collections and multiple tests across markers within collections (Rice 1989). The corrections for multiple tests resulted in low power to detect significant departures from H-W, so we also examined the number of departures from $\mathrm{H}-\mathrm{W}$ by marker and by population prior to correcting for multiple tests to assess any patterns in departures from $\mathrm{H}-\mathrm{W}$.

## Identifying markers under selection

LOSITAN (Antao et al. 2008), an implementation of the FDIST2 package of Beaumont and Nichols (1996), was used to identify markers that produce $\mathrm{F}_{\text {ST }}$ outliers. Markers with high outlier $\mathrm{F}_{\text {ST }}$ values are thought to be under dispersive selection. Due to limitations on the size of dataset used in this program and the geography of the application, we restricted this analysis to populations from the northern Alaska Peninsula, Bristol Bay, and the Kuskokwim River for the 42 nuclear markers. We chose running parameters based upon the following:

1. We chose to not use the "neutral" mean $\mathrm{F}_{\mathrm{ST}}$ setting. This setting estimates a neutral $\mathrm{F}_{\mathrm{ST}}$ from only markers that an initial run of LOSITAN reveals to not be under selection. A second and final run is computed incorporating all markers (giving each an estimated selection status) using the mean neutral $\mathrm{F}_{\mathrm{ST}}$ obtained from the first run described above (Antao et al. 2008 page 3 \# 6). We chose not to use this setting as this simulation analysis suggests that a majority of markers are candidates for balancing selection, more than we believe, and removing this many markers from the estimation of the mean $\mathrm{F}_{\text {ST }}$ results in a spuriously high mean $\mathrm{F}_{S T}$ estimate. However, we ran the analysis both using and not using the 'neutral' setting and found that results do not differ much (e.g., the same markers were identified as candidates for positive selection);
2. We chose to use the force mean $\mathrm{F}_{\mathrm{ST}}$ setting because it approximates the desired average simulated $\mathrm{F}_{\text {ST }}$ to the average value observed in the dataset using a bisection algorithm (Antao et al. 2008 page 3 \# 7);
3. We changed the sample size to more accurately represent the number of individuals we observed in most of the "islands" in our baseline ( $\mathrm{n}=95$ );
4. We removed six populations from the Lake Clark and Upper Kuskokwim regions from the analysis because simulations based upon the island model may not be appropriate for a baseline with these populations included. There is evidence that Lake Clark sockeye salmon populations were recently founded and show signs of a bottleneck effect (Habicht et al. 2004), and there are probably high levels of isolation-by-distance for both of these groups of populations. We chose to remove
these specific populations as they were the most divergent on a Neighbor-Joining tree of pair-wise $\mathrm{F}_{\mathrm{ST}}$ 's (data not shown);
5. We changed the expected number of populations to equal what we included in the simulations (i.e. 90 instead of 96 );
6. We removed five markers from the analysis as they exhibit very low levels of heterozygosity. Beaumont and Nichols recommend discarding markers with heterozygosities less than $2 /$ (sample size), so we used 0.02 as our cut off for removal, which included: One_ctgf-301, One_MARCKS-241, One_p53-534, One_RAG1-103, and One_RH2op-395.

## Population structure visualization

To visualize genetic population structure, Cavalli-Sforza and Edwards (1967) chord distances (CSE) were calculated from allele frequencies at the 42 SNP loci and plotted using the UPGMA method. We chose this measure of genetic distance because previous analyses have identified loci under positive selection and utilizing distance measures that assume neutral loci and are based upon genetic drift (i.e., pair-wise $\mathrm{F}_{\mathrm{ST}}$ 's) may not be appropriate. While this measure is biased by unequal sample sizes, a substantial portion of the populations included in this baseline are of 95 individuals. CSE distances were used to produce two UPGMA trees: 1) all baseline populations and 2) restricted to populations from Western Alaska and the Alaska Peninsula (WAAP).

## Hierarchical log-likelihood analysis

We examined the homogeneity of allele frequencies among populations within regions using a hierarchical log-likelihood ratio test ( $G$ test; Sokal and Rohlf, 1995). We included data from only nuclear loci and removed One_MHC2_251 so as not to duplicate the divergence information provided by the two linked MHC loci. We examined $G$-statistics for each of 17 coastwide regions (Table 1), and summed $G$-statistics and degrees of freedom from 12 of these regions into three broad-scale regions (i.e., Western Bristol Bay YK, Eastern Bristol Bay, and Alaska Peninsula) for an examination of broad-scale population structure. These two levels of analysis
correspond to the regional groupings used in the two UPGMA trees described above. We further summed test statistics across regions into Western Alaska (Norton Sound to South Alaska Peninsula) and Coastwide totals. Finally, we summed test statistics across loci for an overall measure of allele frequency homogeneity at the same hierarchical levels described above. As the number of populations within regions differed greatly (i.e., 3 populations in the Norton Sound region, 116 populations in the Western Gulf of Alaska region), we divided $G$-statistics by degrees of freedom to examine a measure of regional diversity less biased by sampling effort.

## Baseline evaluation for MSA

Reporting groups were delineated based on geographic regions that were thought to be applicable for MSA analyses of mixtures captured under the WASSIP program. Within Norton Sound, Yukon and Kuskokwim Rivers, Bristol Bay and Alaska Peninsula, the reporting groups represent smaller geographic areas on the scale of commercial fishing districts. Outside of these areas, the reporting groups represent much larger geographic areas on the order of management regions or countries. During estimation of stock composition, populations were maintained separately within these reporting groups as recommended by Wood et al. (1987). Reporting group estimates were calculated by summing population estimates.

We then assessed the potential of the baseline to identify these reporting groups for MSA applications with simulations and proof tests. For the simulations, we generated 400 fish based on the population-specific allele frequencies from all the populations within each reporting group (i.e., $100 \%$ simulations). This process was repeated 1,000 times, and the mean and central $90 \%$ of the distribution of estimates were reported as the estimate and the $90 \%$ confidence interval. Simulated mixtures were analyzed using SPAM version 3.7b (Debevec et al. 2000; ADF\&G 2001). For the proof tests, we created a test mixture by sampling approximately 200 fish from each reporting group; we rebuilt the baseline excluding the sampled fish. The test mixture was analyzed using BAYES (Pella and Masuda 2001) with a flat prior (with a weight of one fish). Estimates and $90 \%$ credibility intervals from three chains with different starting conditions were tabulated. We repeated this procedure for each reporting group. For both the simulations and

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proof tests, a critical level of $90 \%$ correct allocation was used to determine if the reporting group was acceptably identifiable (e.g., Seeb et al. 2000).

## Results

## Tissue Sampling

A total of 49,809 individuals from 562 collections representing 375 populations (Table 1; Figure 1) have been genotyped at the 45 SNP markers (Table 2). This baseline represents an increase of 120 populations to the 255 population baseline presented by the ADF\&G Gene Conservation Laboratory (GCL) in its proposal to AYK SSI for WASSIP funding in 2007. Collection sites ranged from the western Kamchatka Peninsula (Russia) to Puget Sound, Washington. The most comprehensive collection was done in the densest portion of the species range, i.e., populations from rivers draining into the Bering Sea and areas adjacent to the Bering Sea (Figure 1). For some analyses we included a subset of collections from the Western Alaska/Alaska Peninsula region (WAAP). This subset was comprised of 20,856 individuals from 221 collections representing 137 populations ranging from the Norton Sound region in the north to the South Peninsula region to the south (Table 1; Figure 2).

## Laboratory Analysis

The overall failure rate for successfully assaying genotypes at the 45 SNP markers in a representative project was $2.3 \%$. The quality control process demonstrated a discrepancy rate of $0.58 \%$. Assuming an equal error rate in the original and quality control genotyping process, our baseline collections were genotyped with a process that produced genotypes with an error rate of $0.29 \%$. An average of 1.4 fish per collection was removed based upon the $80 \%$ rule for the collections that were included in this baseline ( $\mathrm{SD}=3.3$ ). A majority of collections had no fish removed based upon the $80 \%$ rule (i.e., 317), and 102 collections had one fish removed while 12 collections each had greater than 10 fish removed.

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## Statistical Analysis

## Heterozygosity and $F_{S T}$

Observed heterozygosity, expected heterozygosity, and $\mathrm{F}_{\text {ST }}$ for each of the nuclear markers, and only $\mathrm{F}_{\text {ST }}$ for each of the combined loci (see linkage disequilibrium results) are in shown in Table 3. Observed heterozygosity was lower than expected heterozygosity at every nuclear marker with the averages of 0.243 and 0.288 , respectively. Observed heterozygosities ranged widely from 0.017 to 0.447 .

The $\mathrm{F}_{\text {ST }}$ estimate over all markers was 0.149 , but a few nuclear markers had considerably higher values. $\mathrm{F}_{\text {ST }}$ estimates for One_MHC2_251 and $O n e \_M H C 2 \_190 ~ w e r e ~^{0} .303$ and 0.356 , respectively. Other markers with $\mathrm{F}_{\mathrm{ST}}$ estimates greater than 0.2 included: One_Tf_ex10-750, One_HpaI-99, One_STC-410, One_zP3b-49, One_Tf_ex3-182, and One_GHII-2465. The remaining markers had $\mathrm{F}_{\text {ST }}$ values below 0.170 and only three markers had values below 0.050.

## Linkage disequilibrium

Significant gametic disequilibrium was found between one pair of nuclear SNP markers (One_MHC2_190 and One_MHC2_251; Table 4). Other pairs of markers that exhibited linkage disequilibrium within some collections, but below the threshold of $50 \%$ of the populations were: One_GPDH and One_GPDH2 (34\% of collections); One_Tf_ex10-750 and One_Tf_ex3-182 (19\%); and One_RF-112 and One_RF-295 (7\%). All of these pairs are known to be physically linked.

For the pair of linked nuclear SNP markers and the triplet of mitochondrial SNP markers (One_CO1, One_Cytb_17, and One_Cytb_26), genotypes from each marker were pooled to form one haplotype locus: One_MHC2_190_251 and One_CO1_Cytb17_26, respectively. After combining the pair of linked nuclear markers and the three mtDNA markers, the final analyses included 41 independent nuclear loci and 1 mitochondrial locus (described by three SNPs).

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## Pooling collections into populations

The 562 collections reduced to a total of 375 unique populations after pooling collections taken from similar locations over multiple years and from nearby sites that exhibited genetic homogeneity. Some tests for homogeneity between collections within the WAAP area were significant based upon our criterion. Of these, we pooled the following populations with temporal collections based upon the recommendations of Waples (1990): Goodnews River North Fork, Goodnews River Middle Fork, Tommy Creek, Upper Talarik Creek, and Idavain Creek. These represent $18 \%$ of the 28 pairs of collections taken from similar locations over multiple years within the WAAP area. The test for homogeneity between the two collections from the West Fork of the Black River (Chignik drainage) was also significant, but we have little metadata associated with the 1997 collection and so did not pool these collections for this baseline analysis. Technical Document 2 provides a more detailed investigation of this temporal diversity.

The average sample size per population was 133 fish, although a few populations outside the Western Alaska/Alaska Peninsula (WAAP) area were small with as few as 10 fish. Within the WAAP, the smallest population sample size was 47 fish. These populations with sample sizes below 80 fish were mistakenly included in subsequent analyses and are indicated by an asterisk in the population column of Table 1; they will be excluded in the final baseline. A substantial portion of the populations included in this baseline are of 95 individuals (i.e., 115), and 175 populations have a sample size greater than 95 individuals.

## Hardy-Weinberg equilibrium

Significant departures from H-W were not found in any populations for the 42 nuclear SNP markers after correcting for multiple tests. However, before correcting for multiple tests, we did find some patterns in the distribution of departures from H-W. One_MHC2_190 and One_MHC2_251 were out of H-W in 29 and 30 populations, respectively, while no other marker

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was out of $\mathrm{H}-\mathrm{W}$ equilibrium at more than 23 populations (Table 2; Figure 3). Nineteen populations were expected to be out of $\mathrm{H}-\mathrm{W}$ equilibrium for each marker by chance at $\alpha=0.05$.

We also detected eight populations with greater than twice as many markers out of H-W equilibrium than would be expected by chance (before correcting for multiple tests; Table 1 ; Figure 4). Two markers were expected to be out of H-W equilibrium for each population by chance at $\alpha=0.05$. These included Avacha Bay, Dvu 'Yurta River, and Belaia River in Russia, the middle fork of the Goodnews River in western Alaska, Fish Creek and English Bay in Cook Inlet, Mill Creek in southeast Alaska, and Baker Lake in Washington. In all but one of the 61 cases, the significant departure from $\mathrm{H}-\mathrm{W}$ at markers for these populations was due to an excess of homozygotes (i.e., positive $\mathrm{F}_{\text {IS }}$ values).

## Identifying markers under selection

The results of the LOSITAN analysis clearly suggest that the two major histocompatibility complex markers (One_MHC2_190 and One_MHC2_251; MHC) are very different from other markers and that statistically they are candidates for positive selection using these simulation parameters (Figure 5). LOSITAN also suggests One_STC-410 and One_ZNF-61 as candidates for positive selection, although the $\mathrm{F}_{\mathrm{ST}}$ estimate for One_ZNF-61 is not much greater than the upper bound of the mean $\mathrm{F}_{\text {ST }}$ estimate. We would expect 37 (total markers analyzed) minus 2 $($ MHC markers $)=35 \mathrm{X} 0.05$ (alpha) $=2$ markers to be outside the bounds by chance, so excluding candidates for balancing selection, having two markers above the upper bounds is not unreasonable.

The LOSITAN output shows a lower bound that defines many markers as candidates for balancing selection. After removal of the two MHC markers, the $\mathrm{F}_{\mathrm{ST}}$ mean and confidence interval bounds decreased and nine fewer markers are considered candidates for balancing selection. This also then includes two more markers (One_STR07 and One_Prl2) as candidates for positive selection, but these were just above the upper bound (data not shown).

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## Population structure visualization

Genetic relationships among baseline populations are shown schematically in the UPGMA trees (Figures 6 and 7). On the tree with the whole Pacific Rim baseline, the deepest structure was found within the Eastern and Western Gulf of Alaska (Figure 6). A regional structuring of populations was the most common pattern with populations clustered by lakes and drainages. These patterns can most easily be visualized in the WAAP UPGMA (Figure 7), where most of the populations within some of the drainages or nursery lakes cluster together including the Naknek River, Alagnak River, and Chignik River.

Population relationships within some drainages are more complicated than others, which may be the result of a more complicated geography and other factors. The populations within the Wood River, which is made up of five large lakes, beach and tributary spawners, and early- and late-run timing, divide into four clusters. The populations within the Nushagak River, which is a long river with one branch that drains large lakes and other branches that are devoid of lakes, are divided into two clusters and an outlying population. The populations within the Kvichak River, which is made up of one large lake and one smaller lake, are in three clusters with one outlying population. These clusters are made of populations from Lake Clark (highly divergent), northeastern and southwestern Iliamna Lake, and a population spawning between the two lakes. Many of the populations within the North and South Peninsula, which contain many short rivers that drain directly into the ocean, are highly divergent from each other and may reflect the stronger influence of genetic drift on these smaller populations. The populations within the Egegik River cluster into one group representing tributary spawners from the eastern and north side of the nursery lake and a divergent population representing the south side of the nursery lake.

Finally, the Kuskokwim River and Norton Sound contained some of the most divergent collections. These included the Necons River and Telaquana Lake from the Kuskokwim River and Salmon and Glacial lakes that drain into Norton Sound. These Kuskokwim populations and the highly divergent Lake Clark populations were the populations removed from the LOSITAN
analysis for markers under selection and are the most divergent in the WAAP area (top nodes; Figure 7).

## Hierarchical log-likelihood analysis

Substantial heterogeneity in allele frequencies existed among populations within all fine- and broad-scale regions (Table 5). Each test for homogeneity of allele frequencies among populations within regions was highly significant ( $P<0.01$ ). The measure of regional diversity corrected for number of populations (i.e., $G / \mathrm{df}$ ) highlights substantial diversity within particular regions, notably Norton Sound, Yukon Kuskokwim and Kvichak in the WAAP area ( $G / \mathrm{df}=$ 17.27, 18.74, and 21.74, respectively; Figure 8), and Western Gulf of Alaska and Eastern Gulf of Alaska in the coastwide analysis ( $G / \mathrm{df}=37.74$, and 26.16, respectively; Figure 9). Also notable is the relatively low within-region diversity for the WAAP area, especially within the Igushik, Wood, Naknek and Ugashik regions.

Different markers exhibit varying degrees of allele frequency divergence across regions. The One_MHC2_251 marker is the most powerful included in this analysis at describing differences among populations for both the coastwide and WAAP regional scales, and exhibits similar discriminatory power in both regional areas (i.e., $G / \mathrm{df}=82.14$ and 78.88 , respectively). Other markers are very useful at describing coastwide genetic diversity but not as useful within the WAAP study area (e.g., One_E2 G/df $=25.79$ and 9.79, respectively; Figure 10). Similarly, some markers show no differences among populations within some regions (e.g., One_p53-576 $G / \mathrm{df}=0.00$ for Western Kamchatka through Yukon Kuskokwim, data not shown), but very high levels of diversity among populations for other regions (One_p53-576 G / df= 26.36 for Western Gulf of Alaska).

## Baseline evaluation for MSA

Three reporting groups failed to meet the critical level of $90 \%$ correct allocation in the $100 \%$ simulations (Igushik, Ugashik, and North Peninsula; 86\%, 86\% and 89\%, respectively; Figure 11; Table 6). When fish were misallocated in the Igushik simulations, $10 \%$ were allocated to the

Wood River reporting group and $2 \%$ to the Nushagak reporting group. When fish were misallocated in the Ugashik simulations, $4 \%$ were allocated to the Egegik reporting group, 3\% to the North Peninsula reporting group, and $2 \%$ to the Western Gulf of Alaska reporting group. When fish were misallocated in the North Peninsula simulations, $4 \%$ were allocated to the Western Gulf of Alaska reporting group and $2 \%$ to the South Peninsula reporting group. In general, the simulations indicated that most reporting groups can be distinguished from one another with a high degree of accuracy (mean $=93 \%$ ).

Proof tests using the current baseline indicate that the 17 coastwide reporting groups can be distinguished from each other with a high degree of accuracy (mean $=97 \%$; Figure 12; Table 7). Only one of the reporting groups (Western Gulf of Alaska; 89\%) did not meet the critical level of $90 \%$ correct allocation. When fish were misallocated in the Western Gulf of Alaska proof test, $9 \%$ were allocated to the Eastern Gulf of Alaska reporting group.

## Discussion

This sockeye salmon baseline is the most comprehensive SNP database available for any Pacific salmonid. It is also the most comprehensive genetic baseline of any marker type that includes high representation from all areas that are most likely to contribute to mixtures sampled under the WASSIP, with 127 populations from the WAAP areas. The WAAP is also the area where the majority of sockeye salmon are produced. Almost $50 \%$ of all of the sockeye salmon production in the world originate from Bristol Bay drainages alone (Eggers and Irvine 2007; Bugaev et al. 2008). The baseline is least complete for the US/Canada trans-boundary rivers that drain into Southeast Alaska and spawning areas in British Columbia. Major ancestral lineages from those regions that were identified in Beacham et al. (2006) are represented by one or more collections. Thus, despite some gaps in the baseline in this area, adequate samples exist so that fish originating from Eastern Gulf of Alaska populations not included in the baseline will most likely allocate to the large-scale Eastern Gulf of Alaska reporting group.

Population structure for sockeye salmon spanning the Pacific Rim was first described by Beacham et al. (2006). The baseline data for these studies are least complete in the densest
portion of the species range. Such a baseline bias may impact MSA allocations. Their data, for example, indicated that $7 \%$ of a test sample of 62 fish from the western Bering Sea originated from the Alaska Peninsula and none originated from Bristol Bay. Data presented by Habicht et al. (submitted) suggest that Bristol Bay is the dominant regional stock of North American sockeye salmon migrating through the western Bering Sea, and Alaska Peninsula stocks are rarely present. This observation is supported by that of Bugaev et al. (2008), who used scale pattern analysis to report a dominant role for Bristol Bay stocks (55\% of immature sockeye salmon) in summer 2006 BASIS surveys in the REEZ. Nevertheless, Beacham et al. (2006) provide a framework for future studies. The patterns of genetic relationships identified in this study are similar to those reported in Beacham et al. (2006) and provide a template to insure that samples used in this study adequately represent the major lineages of sockeye salmon at the extremes of the species range.

## Marker $F_{S T}$ and resolving power

Beacham et al. (2001) point out that the MHC markers provide a significant portion of the resolving power of the MHC/microsatellite data bases; merging of the MHC portions of the two data sets needs further evaluation given the different analysis methods between the studies. The two MHC markers in our study had the highest $\mathrm{F}_{\text {ST }}$ values among all the markers (Table 3) and the one MHC included in the log-likelihood ratio test had the highest $G$ statistics in both the overall and the WAAP baseline (Figure 10), indicative of the resolving power of this locus for GSI. Among the other markers with high $\mathrm{F}_{\text {ST }}$ values, six others were above 0.2 and included: One_Tf_ex10-750 (0.206); One_HpaI-99(0.218); One_STC-410 (0.220); One_zP3b-49 (0.266); One_Tf_ex3-182 (0.268); and One_GHII-2465 (0.275). Not surprisingly, these six were also identified in the $\log$ likelihood ratio test analysis as the only loci with degree-of-freedomadjusted $G$ statistics higher than 30 for the full baseline (Figure 10).

The log-likelihood ratio test analysis also showed that the loci with the highest $G$ statistics for the full baseline were not identical to those for the WAAP area. For the WAAP area, the $G$ statistics were generally lower with only five loci showing degree-of-freedom-adjusted $G$ statistics above 20. Of these, four of the markers were identified as powerful for discriminating among
populations within regions for the full baseline (the MHC marker, One_Tf_ex10-750; One_HpaI99; and One_zP3b-49), while One_ALDOB-135 was relatively powerful within the WAAP area but intermediate for the full baseline. One_STC_410, One_TFex3-182 and One_GHII-2465 had $G$ statistics below 20 in the WAAP baseline, but higher than 30 in the full baseline. The loglikelihood ratio test might be a good test to identify the most useful markers by region as additional markers become available.

## Markers under selection

Both MHC markers also appeared to be the markers under the strongest positive selection within WAAP (Figure 5). MHC is known to be under selection in salmonids (e.g. Atlantic salmon, Dionne et al. 2007). One_STC-410 was also identified as a candidate locus under selection (Figure 5). One_STC-410 is a SNP for the target locus stanniocalcin, which is a calcium- and phosphate-regulating hormone (Wagner 1994). Some loci with high $\mathrm{F}_{\text {ST }}$ values across the species range were not identified as candidates for positive selection within the WAAP area, but may be under selection outside of this area. These differences in selection and resolving power are indicated as large differences between the measure of within-area diversity ( $G / \mathrm{df}$ ) for the coastwide and WAAP areas in Figure 10 (e.g., One_GHII-2465). One_Zp3b-49 is associated with the zona pellucida, an extracellular matrix that surrounds growing oocytes in mammals and fish and plays a role in gamete recognition, and therefore may be under selection (Epifano et al. 1995). One_Tf_ex10-750 and One_Tf_ex3-182 code for transferrin, which is an iron-binding protein that plays an important role in iron metabolism and resistance to bacterial infection in a variety of organisms. Positive selection for transferrin was detected in an analysis across salmonids (Ford et. al 1999).

The LOSITAN analysis also suggested a large number of markers as candidates for balancing selection. The expected relationships between $\mathrm{H}_{\mathrm{e}}$ and $\mathrm{F}_{\mathrm{ST}}$ were highly affected by the parameters used and the markers included the program. Given the large number of markers that were identified as candidates for balancing selection, more work needs to be done to determine if they are indeed under balancing selection or if some of the model assumptions have been violated. In that effort we are investigating an analysis of these markers in a Bayesian framework (i.e.,

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BayeScan; Foll and Gaggiotti 2008) that may help better identify candidate markers under selection.

## Deviations from H-W

We identified some factors that may explain why some populations were out of $\mathrm{H}-\mathrm{W}$ equilibrium at more than twice the expected number of markers (5 at $P=0.05$, not adjusted for multiple tests). Two of the populations that met this criterion were from places where samples taken early and late within calendar years were pooled (English Bay and Mill Creek). When chi-square tests were performed to test for homogeneity among these collections, English Bay had a $P$-value of 0.02 and Mill Creek had a value above 0.05 . These $P$-values were above our critical value of 0.01 for pooling collections into populations. One possibility that either the early or late collections were mixtures of two run timings which resulted in the large number of markers out of $\mathrm{H}-\mathrm{W}$ while producing relatively high $P$-values in the chi-square tests.

Three of the populations out of H-W equilibrium were taken in Russia and we have little metadata to determine which factors may contribute to departures from H-W (Avacha Bay, Dvu 'Yurta River, and Belaia River). The large number of deviant markers for Avacha Bay (12) indicates that this collection may be made up from a combination of populations, separated either temporally or spatially, but we have little information for this collection. The Dvu 'Yurta and Belaia river populations are each combinations of two collections taken in consecutive years. Again we do not have calendar day for these collections or any other metadata, but the $P$-values for the chi-square tests were below 0.01 for both of these tests, indicating that the collections differed between the two years. In future baseline analyses we may want to exclude the 1995 collections because they contain only 11 fish each.

The Middle Fork Goodnews River population was made up of three collections (1991, 2001, and 2007) and the chi-square test was highly significant ( $P<0.01$ ). The 2007 collection was made throughout June and July, while the other collections were made in mid July and early August indicating that there may be multiple populations in these samples that are temporally segregated.

The two Fish Creek collections were taken at similar calendar dates 16 years apart and had a highly significant chi-square test result ( $P<0.01$ ). These collections are of fish captured at the Fish Creek weir and may be a mixture of populations that segregate spatially within the Fish Creek drainage. These collections could not be pooled with the Fish Creek samples taken at the Big Lake Hatchery, which is in the Fish Creek drainage. This year we collected fish in Meadow Creek, another tributary to Fish Creek, with the hope that this collection can substitute for the weir collection in future baselines.

Finally, the collection from Baker Lake had more than five markers out of H-W equilibrium. We have no metadata from this location, but spatially segregated natural and artificial spawning areas that are used in Baker Lake to mitigate for dams (http://wdfw.wa.gov/fish/sockeye/bakerriver.htm) might be becoming reproductively isolated (i.e. Hendry et al. 2000). All but one of these departures from H-W expectations are the result of an excess of homozygotes, indicative of a Wahlund effect and consistent with observing an admixture of populations.

## Population structure

The hierarchical analysis of allele frequency homogeneity highlighted high levels of diversity observed for some regions (e.g., Kvichak, Western Gulf of Alaska and Eastern Gulf of Alaska; Figures 8 and 9), although the range of many of the defined regions was large. These observations are often driven by large differences in allele frequencies observed between large groups of populations or for few outlier populations. Within the Kvichak region, this is the result of a strong divergence between populations within the Lake Clark and Iliamna nursery lakes that has been previously described (Habicht et al. 2004). The Western Gulf of Alaska region encompasses a geographically broad region with high levels of divergence among populations within the region. This divergence is largely driven by the clustering of populations within the Kodiak Archipelago, Kenai, Susitna and Copper rivers (Figure 6). In contrast, the large diversity observed within the Eastern Gulf of Alaska region results from a few highly deviant outlier populations (i.e., Kanalku Lake, Mahoney Creek, Tahltan Lake, Little Tahltan Lake, and Kah

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Sheets Lake) with allele frequencies very discordant from two large, loosely clustered groups of the remaining populations. There is relatively little genetic diversity observed within the WAAP study area compared to the Gulf of Alaska regions, which may be the result of a more recent common ancestral population in the Beringia Refugium and many populations with large population sizes that likely retards the influence of genetic drift on genetic divergence.

Aside from some notable exceptions such as Norton Sound, Upper Kuskokwim and Lake Clark, the WAAP study area shows lower levels of genetic differentiation than areas in the Eastern and Western Gulf of Alaska (Figure 9, Table 5).

## Baseline evaluation

Simulation and proof test results indicate that the 17 coastwide reporting groups can be distinguished from each other with a reasonable degree of accuracy. The two methods differ in that simulations generate hypothetical individuals from baseline allele frequencies, whereas proof tests remove known individuals from the baseline to be treated as mixture individuals. As such the proof tests provide a more realistic and robust methodology for testing the utility of the baseline at discriminating among reporting groups for GSI purposes. When fish were misallocated they were most often allocated to neighboring reporting groups and/or reporting groups with populations with very similar allele frequencies. For example, Pick Creek in the Wood River reporting group has allele frequencies similar to all of the Igushik populations, groups together with Igushik populations on trees, and can cause misallocation between these two adjacent reporting groups.

There are a number of potential sources of improvement in our baseline evaluation tests. The proof tests, for example, included only 200 individuals yet the WASSIP mixtures will generally be made up of 400 fish. The small sample sizes in the proof tests were necessitated by the small sample size of one reporting group (Norton Sound; 335 fish). The inclusion of additional SNPs will also likely increase resolving power due to an increase in the number of independent markers as well as the potential that some of the new SNPs are under selection and may represent adaptive differences among populations in the WASSIP area. Baseline evaluations that
are comprised of more heterogeneous mixture compositions (i.e., not $100 \%$ ) will provide a measure of baseline utility at discriminating among reporting groups in a more realistic fashion. There are statistical improvements that may improve our GSI resolving power and the results of baseline evaluation tests. Two such examples are the use of informative priors when using Bayesian methods for GSI and the use of a stratified estimate protocol (Technical Document 3).

## Future analyses

1. Increase sample sizes for collections for which we have existing tissues to be genotyped.
2. Incorporate collections gathered in the 2009 field collection season into baseline analyses.
3. Remove populations with samples sizes of less than 80 fish (denoted with an asterisk in Table 1) for which we do not have existing tissues to be genotyped from the baseline.
4. Investigate temporal variation in allele frequencies for collections from similar locations in multiple years. Is this variation driven by loci under selection? Does this variation represent problems with our genotyping process? We foresee resampling populations to ensure that the baseline data are still valid and to help address these concerns.
5. Assess the suite of developing SNPs (see Technical Document 6) for utility in describing genetic variation within the WASSIP study area and for accurately and precisely estimating stock proportions in mixture samples from area fisheries.
6. Perform proof tests with 400 fish in reporting groups where adequate numbers of fish exist.
7. Perform simulations and proof tests using more heterogeneous mixture compositions (i.e., not $100 \%$ ) to assess baseline utility at discriminating among reporting groups in a more realistic fashion.
8. Investigate why we saw a consistent pattern of lower observed heterozygosities than expected (Table 3).
9. Further investigate the utility of the loci identified in LOSITAN as loci under balancing selection. Loci under balancing selection may be good candidates to be replaced with loci under positive selection for MSA as new markers become available.
10. Conduct further analyses of genetic diversity, including AMOVA and Nei's gene diversity analysis, and examine $G$ statistics for hierarchical levels within the WAAP area that may have more biologic meaning (e.g., populations within nursery lakes).
11. For these other levels of hierarchy, compare levels of heterogeneity using Fisher's $F$-test to better understand how diversity is distributed in the baseline.
12. Examine the distribution of allelic richness by region and ascertainment region to assess ascertainment bias.
13. Utilize statistical methods developed for estimating small proportions to increase the performance of MSA through decreased bias and increased precision. These methods might include the use of informative priors when using Bayesian methods for GSI and the use of a stratified estimate protocol (Technical Document 3)
14. Investigate the utility of reducing the range of the baseline to include only those populations that are likely to be present in WASSIP mixtures.

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## Technical Committee review and comments

## Document 5: Status of the SNP baseline for sockeye salmon

Figure 3. It is worth noting that the null expectation for no linkage disequilibrium implicitly assumes an infinite parental population. One actually expects more than the nominal alpha fraction of significant tests simply due to drift. The fact that no general elevation of significant LD was found, despite rather large samples, suggests that most populations do not have small Ne .

Tests for selection. We also are suspicious of results of programs that suggest large numbers of loci apparently under selection. Evidence is accumulating that methods currently in use to identify 'outlier' loci do not fully account for variance in Fst due to historical population demography and population structure. See in particular the two references below:

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[Unedited comments from "Panel comments October 2009.doc" related to Technical Document 5.

## Tables

Table 1. Baseline collection information organized geographically by reporting group and subdivided by population. Each line contains an individual collection with associated collection name, collection date (only year is provided for collections where calendar day was not known), and sample size. Some collections were pooled based on geographic proximity and tests of homogeneity (see text for methods). Collections that were pooled fall under the same number under the "Pop \#" column. Populations that were out of $\mathrm{H}-\mathrm{W}$ at more than twice the number of loci than expected by chance ( 5 loci @ $P=0.05$ ) are noted with the number of loci out of $\mathrm{H}-\mathrm{W}$ equilibrium under the H-W column. Populations with an asterisk ( ${ }^{*}$ ) were represented by collections with total sample sizes of less than 80 fish. These populations will either have sample sizes increased in subsequent genotyping efforts or be dropped from future analyses.

| Reporting group | Pop \# Population | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: |
| Western Kamchatka | 1 Palana River | Palana River | 6/27/2002 | 48 |
|  |  | Palana River | 2002 | 50 |
|  | 2 Tigil River | Tigil River | 6/18/2002 | 100 |
|  | 3 Bistraya River* | Bistraya River | 8/16/1998 | 56 |
|  | 4 Bolshaya River* | Bolshaya River | 8/16/1999 | 29 |
|  |  | Bolshaya River | 2003 | 40 |
|  | 5 Kuril Lake | Etamink River Early | 8/21/1990 | 29 |
|  |  | Etamink River Late | 9/28/1990 | 48 |
|  |  | Kirushutk River | 2000 | 49 |
|  |  | Etamink River | 8/12/2002 | 46 |
|  |  | Khakizun Bay | 8/25/2002 | 49 |
|  |  | North Far Bay | 8/26/2002 | 50 |
|  | 6 Gabruschka Bay* | Gabruschka Bay | 8/25/2002 | 49 |
|  | 7 Vichenkiya River | Vichenkiya River | 2000 | 96 |
|  | 8 Olada Bay* | Olada Bay | 2000 | 50 |
|  | 9 Ozernaya Bay | Ozernaya Bay | 2000 | 50 |
|  |  | Ozernaya River | 2000 | 49 |
|  |  | Ozernaya River | 8/5/2003 | 50 |

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| Reporting group | Pop \# Population |  | H-W Collection |  | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Ozernaya River | 8/14/2002 | 50 |
|  |  |  |  |  |  | 988 |
| Eastern Kamchatka | 10 | Avacha Bay* | 12 | Avacha Bay |  | 60 |
|  | 11 | Kitilgina River* |  | Kitilgina River | 6/29/1998 | 28 |
|  | 12 | Kozireuka River* |  | Kozireuka River | 1994 | 40 |
|  | 13 | Dvu 'Yurta River | 9 | Dvu 'Yurta River | 1994 | 77 |
|  |  |  |  | Dvu 'Yurta River | 1995 | 11 |
|  | 14 | Belaia River | 7 | Belaia River | 1994 | 69 |
|  |  |  |  | Belaia River | 1995 | 11 |
|  | 15 | Hapiza River |  | Hapiza River Early | 7/17/1998 | 96 |
|  |  |  |  | Hapiza River Late | 9/2/1998 | 79 |
|  | 16 | Elovka River |  | Elovka River | 1994 | 69 |
|  |  |  |  | Elovka River | 1995 | 40 |
|  | 17 | Azabachje Lake* |  | Azabachje Lake | 2004 | 30 |
|  | 18 | Kamchatka River Early* |  | Kamchatka River Early | 6/1/1998 | 79 |
|  | 19 | Kamchatka River Late |  | Kamchatka River Late | 7/21/1998 | 97 |
|  | 20 | Lake Potat* |  | Lake Potat | 7/29/2001 | 49 |
|  | 21 | Lake Vati* |  | Lake Vati | 8/7/2002 | 48 |
|  | 22 | Anana Lagoon* |  | Anana Lagoon Early | 6/24/2002 | 30 |
|  |  |  |  | Anana Lagoon Late | 7/4/2002 | 48 |
|  | 23 | Severnaya Lagoon |  | Severnaya Lagoon | 6/26/2002 | 97 |
|  |  |  |  |  |  | 1,058 |
| Norton Sound | 24 | Salmon Lake |  | Salmon Lake | 8/3/2001 | 96 |
|  | 25 | Glacial Lake |  | Glacial Lake | 8/15/2004 | 144 |
|  | 26 | Unalakleet River |  | Unalakleet River | 8/22/2007 | 95 |
|  |  |  |  |  |  | 335 |

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| Reporting group | Pop \# Population |  |  | Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Yukon Kuskokwim | 27 | Gisasa River* |  | Gisasa River | 7/16/2005 | 47 |
|  |  |  |  | Gisasa River | 6/28/2006 | 18 |
|  | 28 | Andreafsky River |  | Andreafsky River | 6/28/2006 | 48 |
|  |  |  |  | Andreafsky River | 7/19/2008 | 46 |
|  | 29 | Necons River |  | Necons River | 8/1/2006 | 55 |
|  |  |  |  | Necons River | 7/28/2007 | 95 |
|  | 30 | Telaquana Lake Outlet |  | Telaquana Lake Outlet | 8/14/2003 | 96 |
|  | 31 | Telaquana Lake Beach* |  | Telaquana Lake Beach | 10/4/2005 | 47 |
|  | 32 | Kogrukluk River |  | Kogrukluk River | 7/6/2001 | 96 |
|  |  |  |  | Kogrukluk River | 7/24/2007 | 48 |
|  | 33 | Salmon River |  | Salmon River | 8/2/2006 | 142 |
|  | 34 | Kwethluk River |  | Kwethluk River | 2007 | 141 |
|  | 35 | Kanektok River |  | Kanektok River | 7/16/2002 | 95 |
|  |  |  |  | Kanektok River | 7/10/2007 | 48 |
|  | 36 | Goodnews River North Fork |  | Goodnews River North Fork | 7/23/2002 | 95 |
|  |  |  |  | Goodnews River North Fork | 7/20/2006 | 47 |
|  | 37 | Goodnews River Middle Fork | 6 | Goodnews River Middle Fork | 8/1/1991 | 48 |
|  |  |  |  | Goodnews River Middle Fork | 7/15/2001 | 96 |
|  |  |  |  | Goodnews River Middle Fork | 6\&7/2007 | 47 |
|  |  |  |  |  |  | 1,355 |
| Togiak | 38 | Togiak River |  | Togiak Lake, Sunday Creek | 8/21/2000 | 94 |
|  |  |  |  | Togiak Lake, Outlet | 7/27/2006 | 95 |
|  | 39 | Ongivinuk Lake |  | Ongivinuk Lake | 8/24/2006 | 142 |
|  | 40 | Nenevok Lake |  | Nenevok Lake | 8/24/2006 | 142 |
|  | 41 | Gechiak Lake |  | Gechiak Lake | 8/21/2000 | 96 |
|  | 42 | Kulukak Lake |  | Kulukak Lake | 8/24/2006 | 142 |


| Reporting group | Pop \# Population | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 711 |
| Igushik | 43 Ualik Lake | Ualik Lake | 8/14/2003 | 95 |
|  | 44 Ongoke Lake Lower | Ongoke Lake Lower | 8/28/2007 | 143 |
|  | 45 Ongoke Lake Upper | Ongoke Lake Upper | 8/27/2007 | 94 |
|  | 46 Amanka Lake | Amanka Lake | 8/14/2003 | 94 |
|  |  |  |  | 426 |
| Wood | 47 Lake Kulik beaches | Lake Kulik beaches | 9/10/2007 | 95 |
|  |  | Lake Kulik beaches | 9/10/2007 | 78 |
|  |  | Lake Kulik beaches | 7/27/2008 | 8 |
|  | 48 Grant River | Grant River | 8/22/2007 | 92 |
|  | 49 Lake Kulik | Lake Kulik | 8/1/2001 | 96 |
|  | 50 Silver Horn Beaches | Silver Horn Beaches | 9/10/2007 | 95 |
|  |  | Silver Horn Beaches | 9/10/2007 | 94 |
|  |  | Silver Horn Beaches | 7/27/2008 | 124 |
|  | 51 Hardluck Bay | Hardluck Bay Beaches | 9/10/2007 | 95 |
|  |  | Hardluck Bay | 9/1/2008 | 156 |
|  | 52 Agulukpak River | Agulukpak River | 8/21/2001 | 96 |
|  | 53 Anvil Bay Beach | Anvil Bay Beach | 8/20/2006 | 94 |
|  |  | N4 Beach | 8/11/2006 | 94 |
|  | 54 Little Togiak Lake | A Beach | 8/8/2004 | 65 |
|  |  | A Beach | 8/10/2005 | 30 |
|  | 55 Pick Creek | Pick Creek | 8/3/2001 | 93 |
|  |  | Pick Creek | 7/22/2008 | 90 |
|  | 56 Sixth Creek | Sixth Creek | 8/1/2008 | 94 |
|  | 57 Agulowok River | Agulowok River | 8/22/2001 | 95 |
|  | 58 Lynx Beach | Lynx Beach | 8/11/2006 | 95 |

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| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 59 | Lynx Creek | Lynx Creek | 8/22/2001 | 96 |
|  | 60 | Ice Creek Upper* | Ice Creek Upper | 8/10/2007 | 67 |
|  | 61 | Aleknagik Lake Creeks | Happy Creek | 7/30/2001 | 95 |
|  |  |  | Bear Creek | 8/2/2001 | 96 |
|  |  |  | Hansen Creek | 8/4/2004 | 95 |
|  |  |  | Ice Creek Lower | 8/9/2007 | 95 |
|  | 62 | Yako Creek* | Yako Creek | 8/1/2008 | 68 |
|  | 63 | Yako Beach | Yako Beach | 8/19/2006 | 95 |
|  | 64 | Eagle Creek | Eagle Creek | 8/12/2007 | 93 |
|  | 65 | Mission Creek | Mission Creek | 1998 | 93 |
|  |  |  |  |  | 2,672 |
| Nushagak | 66 | Mulchatna River Upper | Mulchatna River | 8/27/2001 | 96 |
|  |  |  | Mulchatna River | 8/27/2001 | 65 |
|  | 67 | Mulchatna River Lower | Koktuli River | 8/13/2000 | 96 |
|  |  |  | Stuyahok River | 8/14/2000 | 96 |
|  | 68 | Nushagak River Upper | Klutapuk Creek | 8/18/2001 | 95 |
|  |  |  | King Salmon River | 8/18/2001 | 96 |
|  |  |  | Upper Nushagak Sloughs | 8/19/2001 | 96 |
|  | 69 | Chauekuktuli Lake beach | Chauekuktuli Lake Beach | 8/22/2001 | 96 |
|  | 70 | Allen River | Allen River | 8/22/2001 | 95 |
|  | 71 | Allen River Beach | Allen River Beach | 8/17/2000 | 95 |
|  | 72 | Nuyakuk Lake | Nuyakuk Lake | 8/16/2000 | 99 |
|  |  |  | Nuyakuk Lake South Beach | 8/23/2001 | 94 |
|  | 73 | Tikchik Lake Creek | Tikchik Lake Creek | 8/18/2000 | 95 |
|  | 74 | Tikchik River | Tikchik River | 8/18/2001 | 96 |
|  |  |  |  |  | 1,310 |

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| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Kvichak | 75 | Tlikakila River | Tlikakila River Glacier Fork | 10/6/1999 | 47 |
|  |  |  | Tlikakila River Upper | 9/24/2001 | 96 |
|  | 7677 | Little Lake Clark | Little Lake Clark | 10/9/1999 | 95 |
|  |  | Kijik River Lower | Kijik River Lower | 9/18/2001 | 96 |
|  | 77 78 | Kijik River | Kijik River | 9/19/2001 | 96 |
|  | 7879 | Chulitna Lodge Beach | Chulitna Lodge Beach | 10/5/1999 | 100 |
|  |  |  | Chulitna Lodge Ponds | 10/1/1999 | 47 |
|  | 80 | Sucker Bay Lake | Sucker Bay Lake | 9/14/2007 | 95 |
|  | 81 | Newhalen River | Tazimina River | 8/29/2001 | 96 |
|  |  |  | Newhalen River | 9/3/2002 | 96 |
|  | 82 | Tomkok Creek | Tomkok Creek | 8/24/2000 | 95 |
|  |  |  | Tomkok Creek | 8/28/2002 | 48 |
|  | 83 | Northeast Iliamna Lake | Knutson Bay Late | 10/16/1999 | 95 |
|  |  |  | Bear Pond Late | 10/17/1999 | 47 |
|  |  |  | Grass Pond Late | 10/15/1999 | 44 |
|  |  |  | Pedro Ponds | 1999 | 47 |
|  |  |  | Knutson Bay | 8/27/2000 | 96 |
|  | 84 | East Iliamna Lake | Chinkelyes Creek | 8/28/2000 | 97 |
|  |  |  | Finger Beach 1 | 8/24/2000 | 84 |
|  |  |  | Iliamna River | 8/21/2004 | 46 |
|  | 85 | Iliamna River Late | Iliamna River Late | 10/17/1999 | 96 |
|  | 86 | Iliamna Lake Islands | Fuel Dump Island | 8/28/2000 | 99 |
|  |  |  | Triangle Island | 8/16/2000 | 96 |
|  |  |  | Woody Island West Beach | 8/19/2001 | 100 |
|  | 87 | Tommy Creek | Tommy Creek | 8/24/2000 | 96 |
|  |  |  | Tommy Creek | 8/19/2002 | 48 |
|  | 88 | Copper River | Copper River | 8/23/1999 | 47 |
|  |  |  | Copper River | 8/28/2000 | 96 |

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| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 89 | South Iliamna Lake | Gibralter River | 8/23/1999 | 47 |
|  |  |  | Belinda Creek | 8/25/2000 | 95 |
|  |  |  | Dennis Creek | 8/23/2000 | 96 |
|  |  |  | Gibralter River | 8/25/2000 | 100 |
|  |  |  | Nick N Creek | 8/25/2000 | 96 |
|  | 90 | Gibraltar Lake | Southeast Creek | 8/26/2000 | 96 |
|  |  |  | Dream Creek | 8/22/2001 | 96 |
|  | 91 | Upper Talarik Creek | Upper Talarik Creek | 8/15/2004 | 94 |
|  |  |  | Upper Talarik Creek | 8/10/2006 | 94 |
|  | 92 | Lower Talarik Creek | Lower Talarik Creek | 8/26/2000 | 96 |
|  |  |  | Lower Talarik Creek | 8/23/2001 | 70 |
|  |  |  |  |  | 3,221 |
| Alagnak | 93 | Moraine Creek | Moraine Creek | 9/4/2001 | 96 |
|  |  |  | Funnel Creek Early | 8/8/2004 | 171 |
|  |  |  | Moraine Creek | 9/9/2004 | 96 |
|  |  |  | Moraine Creek Early | 8/8/2004 | 190 |
|  | 94 | Battle Lake | Battle Creek | 9/4/2001 | 96 |
|  |  |  | Battle Creek | 9/8/2004 | 96 |
|  |  |  | Battle Lake Beach | 9/11/2004 | 190 |
|  |  |  | Battle Lake Tributary | 9/11/2004 | 192 |
|  | 95 | Nanuktuk Creek | Nanuktuk Creek | 9/9/2004 | 191 |
|  |  |  | Nanuktuk Creek Early | 8/9/2004 | 190 |
|  | 96 | Kulik River | Kulik River | 9/5/2001 | 96 |
|  |  |  | Kulik River | 9/8/2004 | 96 |
|  |  |  |  |  | 1,700 |
| Naknek | 97 | American River | American River | 8/22/2000 | 95 |


| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | American River | 8/17/2001 | 95 |
|  | 98 | Grosvenor Lake | Grosvenor Lake | 8/12/2003 | 96 |
|  | 99 | Hardscrabble Creek | Hardscrabble Creek | 8/12/2003 | 95 |
|  | 100 | Iliuk Arm | Katolinat Creek \#1 | 9/17/2006 | 48 |
|  |  |  | Margot Creek | 8/15/2001 | 95 |
|  | 101 | East La Gorce Creek* | East La Gorce Creek | 8/27/2006 | 47 |
|  | 102 | Headwater Creek | Headwater Creek | 7/22/2001 | 132 |
|  | 103 | Brooks Lake | Brooks Lake | 8/22/2000 | 100 |
|  | 104 | Dumpling Creek \#1* | Dumpling Creek \#1 | 8/26/2006 | 48 |
|  | 105 | Dumpling Creek \#3 | Dumpling Creek \#3 | 9/17/2006 | 83 |
|  | 106 | Charlene Creek* | Charlene Creek | 9/11/2006 | 47 |
|  | 107 | Lower Q-Tip Lake | Lower Q-Tip Lake | 9/12/2006 | 86 |
|  | 108 | North La Gorce Creek* | North La Gorce Creek | 9/10/2006 | 47 |
|  | 109 | Idavain Creek | Idavain Creek | 8/23/2000 | 96 |
|  |  |  | Idavain Creek | 8/29/2006 | 48 |
|  |  |  |  |  | 1,258 |
| Egegik | 110 | East Becharof Lake | Becharof Creek | 8/11/2000 | 96 |
|  |  |  | Cabin Creek | 8/15/2000 | 96 |
|  |  |  | Ruth Lake Outlet | 8/12/2000 | 95 |
|  |  |  | Cleo Creek | 8/16/2001 | 95 |
|  |  |  | Featherly Creek | 8/16/2001 | 95 |
|  |  |  | Burls Creek | 8/16/2006 | 93 |
|  |  |  | Salmon Creek | 8/16/2006 | 190 |
|  | 111 | Kejulik River | Kejulik River Upper | 8/8/2000 | 47 |
|  |  |  | Kejulik River | 8/17/2001 | 96 |
|  | 112 | Becharof Lake North | Becharof Lake North Tributary | 8/11/2008 | 189 |
|  | 113 | Becharof Lake South | Becharof Lake South Beach | 8/11/2008 | 189 |

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| Reporting group | Pop \# Population | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 1,281 |
| Ugashik | 114 Ugashik Creek | Ugashik Creek | 7/21/2001 | 96 |
|  | 115 Ugashik Lake | Ugashik Narrows | 8/24/2000 | 97 |
|  |  | Deer Creek | 7/20/2001 | 96 |
|  |  | East Creek Mouth | 8/8/2005 | 95 |
|  |  | Black Creek | 8/24/2005 | 95 |
|  | 116 Outlet Stream | Outlet Stream | 8/26/2000 | 96 |
|  | 117 Figure 8 Creek | Figure 8 Creek | 8/22/2005 | 94 |
|  | 118 Old Ham Creek | Old Ham Creek | 8/22/2005 | 95 |
|  |  |  |  | 764 |
| North Peninsula | 119 Cinder River | Mainstem Cinder River | 7/29/2005 | 95 |
|  |  | Wiggly Creek | 7/29/2005 | 80 |
|  | 120 Lava Creek | Lava Creek | 7/23/2004 | 92 |
|  |  | Mud Creek A | 7/30/2005 | 95 |
|  | 121 Meshik Lake | Meshik Lake Shoals | 7/30/2005 | 95 |
|  |  | Meshik Lake Outlet | 7/30/2005 | 95 |
|  | 122 Meshik River | Blue Violet Creek | 7/29/2002 | 92 |
|  |  | Landlock Creek | 7/29/2002 | 96 |
|  |  | L Creek | 7/30/2005 | 95 |
|  | 123 Red Bluff Creek | Red Bluff Creek | 7/30/2005 | 95 |
|  | 124 Willie Creek | Willie Creek | 8/27/2001 | 81 |
|  | 125 Wildman Lake | Wildman Lake | 7/30/2005 | 94 |
|  | 126 Ocean River | Ocean River | 2001 | 96 |
|  | 127 Sandy Lake | Sandy Lake | 6/30/2000 | 96 |
|  |  | Sandy Lake | 7/8/2007 | 95 |
|  | 128 Bear River Early | Bear River Early | 6/30/2000 | 96 |

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| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 129 | Bear River Late | Bear River Late | 8/18/2000 | 96 |
|  | 130 | Hoodoo Lake | Hoodoo Lake | 7/31/2001 | 95 |
|  |  |  | Hoodoo Lake Shoals | 7/31/2005 | 95 |
|  |  |  | Nelson River | 2007 | 47 |
|  | 131 | Nelson River | Nelson River | 7/5/2000 | 96 |
|  | 132 | Davids River | Davids River | 7/31/2005 | 95 |
|  | 133 | North Creek | North Creek | 7/25/2007 | 91 |
|  | 134 | Paul Hansen Tributary | Paul Hansen Tributary | 7/30/2002 | 95 |
|  | 135 | Outer Marker Lake | Outer Marker Lake | 9/9/2004 | 95 |
|  | 136 | Swanson's Lagoon | Swanson's Lagoon | 8/25/2008 | 95 |
|  | 137 | Peterson Lagoon | Peterson Lagoon | 8/2/2005 | 95 |
|  | 138 | Whaleback Mountain Creek | Whaleback Mountain Creek | 7/30/2002 | 96 |
|  | 139 | Summer Bay Lake | Summer Bay Lake | 8/25/1999 | 96 |
|  | 140 | McLees Lake | McLees Lake | 6/4/2004 | 142 |
|  |  |  |  |  | 2,817 |
| South Peninsula | 141 | Hansen Lake | Hansen Lake | 8/2/2005 | 95 |
|  | 142 | Middle Lagoon | Middle Lagoon | 7/28/2004 | 142 |
|  | 143 | Thin Point Lagoon | Thin Point Lagoon | 8/1/2005 | 95 |
|  | 144 | Mortensen's Lagoon | Mortensen's Lagoon | 8/2/2004 | 142 |
|  | 145 | Long John Lagoon | Long John Lagoon | 8/1/2005 | 95 |
|  | 146 | Archeredin Lake | Archeredin Lake | 8/3/2005 | 95 |
|  | 147 | Sanak Island | Sanak Island | 8/24/2008 | 86 |
|  | 148 | Canoe Bay River | Canoe Bay River | 8/26/2008 | 95 |
|  | 149 | Orzinski | Orzinski | 7/1/2000 | 95 |
|  | 150 | Black Lake | Big Spring | 1997 | 95 |
|  |  |  | Broad Creek | 9/1/1997 | 94 |
|  |  |  | Boulevard Creek | 9/1/1997 | 95 |

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| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Tern Lake | 1993 | 95 |
|  | 237 | Quartz Creek | Quartz Creek | 8/6/1993 | 95 |
|  | 238 | Between Skilak and Kenai Lakes | Russian River below falls | 8/2/1993 | 93 |
|  |  |  | Kenai River Late | 9/11/1993 | 47 |
|  |  |  | Kenai River Early | 8/18/1993 | 48 |
|  |  |  | Kenai River Site 1 | 8/22/1994 | 47 |
|  |  |  | Kenai River Site 2 | 8/22/1994 | 48 |
|  |  |  | Kenai River Site 4 | 8/22/1994 | 48 |
|  |  |  | Kenai River Early | 1994 | 96 |
|  |  |  | Kenai River Site 3 | 8/22/1994 | 47 |
|  |  |  | Kenai River Site 5 | 9/9/1994 | 95 |
|  | 239 | Upper Russian Lake Late Bear Creek | Upper Russian Lake Late Bear Creek | 8/29/1997 | 94 |
|  | 240 | Upper Russian Lake Early | Upper Russian River Early, Weir | 7/1/1992 | 96 |
|  |  |  | Goat Creek | 8/19/1997 | 95 |
|  | 241 | Upper Russian Lake Late South | Upper Russian Lake Late South | 9/16/1999 | 95 |
|  | 242 | Upper Russian Lake Late North | Upper Russian Lake Late North | 9/17/1999 | 95 |
|  | 243 | Lower Russian Lake Late Outlet | Lower Russian Lake Late Outlet | 8/2/1993 | 95 |
|  | 244 | Hidden Lake | Hidden Creek | 7/29/1993 | 95 |
|  |  |  | Hidden Lake North Shore | 9/23/2008 | 95 |
|  | 245 | Skilak Lake Outlet | Skilak Lake | 8/1/1992 | 96 |
|  |  |  | Skilak Lake Outlet Early | 1994 | 140 |
|  |  |  | Skilak Lake Outlet Late | 1994 | 140 |
|  |  |  | Skilak Lake | 1995 | 48 |
|  | 246 | Tustumena Lake | Moose Creek | 8/1/1992 | 96 |
|  |  |  | Nikolai Creek | 7/1/1992 | 95 |
|  |  |  | Bear Creek | 8/10/1993 | 95 |
|  |  |  | Glacier Flats Creek | 8/4/1994 | 95 |
|  |  |  | Seepage Creek | 1994 | 95 |



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| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 268 | Klutina River | Klutina River | 8/21/2008 | 156 |
|  | 269 | Long Lake | Long Lake | 9/7/2005 | 95 |
|  | 270 | Tebay River | Tebay River | 8/18/2008 | 197 |
|  | 271 | Bremner River Salmon Creek | Bremner River Salmon Creek | 8/17/2008 | 99 |
|  | 272 | Bremner River Steamboat Lake | Bremner River Steamboat Lake | 8/17/2008 | 177 |
|  | 273 | Clear Creek | Clear Creek | 8/24/2007 | 94 |
|  | 274 | Martin Lake | Martin Lake | 7/26/2007 | 95 |
|  | 275 | Kushtaka Lake | Kushtaka Lake | 8/9/2007 | 95 |
|  | 276 | Bering Lake | Bering Lake | 7/12/1991 | 95 |
|  |  |  |  |  | 17,259 |
| Eastern GOA | 277 | East Alsek River | East Alsek River | 10/15/2000 | 96 |
|  | 278 | Klukshu River | Klukshu River | 8/23/2006 | 95 |
|  | 279 | Upper Tatshenshini | Upper Tatshenshini | 2003 | 95 |
|  | 280 | Neva Lake | Neva Lake | 7/11/2008 | 94 |
|  | 281 | Chilkat River Bear Flats | Chilkat River Bear Flats | 8/9/2007 | 95 |
|  | 282 | Chilkat River Mule Meadows | Chilkat River Mule Meadows | 8/1/2003 | 95 |
|  | 283 | Chilkat River Mosquito Lake | Chilkat River Mosquito Lake | 8/4/2007 | 95 |
|  | 284 | Chilkat Lake Early | Chilkat Lake Early | 7/29/2007 | 95 |
|  | 285 | Chilkat Lake Late | Chilkat Lake Late | 8/12/2007 | 95 |
|  | 286 | Chilkoot River | Chilkoot River | 10/3/2003 | 95 |
|  | 287 | Chilkoot Lake Beaches | Chilkoot Lake Beaches | 7/21/2007 | 95 |
|  | 288 | Berners Bay | Berners Bay | 8/18/2003 | 95 |
|  | 289 | Windfall Lake | Windfall Lake | 7/31/2003 | 48 |
|  |  |  | Windfall Lake | 8/2/2007 | 48 |
|  | 290 | Steep Creek | Steep Creek | 8/20/2003 | 95 |
|  | 291 | Nahlin River | Nahlin River | 7/31/2003 | 50 |
|  |  |  | Nahlin River | 7/31/2007 | 34 |


| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 292 | Tatsamenie Lake | Tatsamenie Lake | 1992 | 95 |  |
|  | 293 | Tatsamenie Lake | Tatsamenie Lake | 2005 | 95 |
|  | 294 | Little Tatsamenie Lake | Little Tatsamenie Lake | $9 / 21 / 1990$ | 64 |
|  | 295 | Little Trapper Lake | Little Tatsamenie Lake | $9 / 11 / 1991$ | 25 |
|  | 296 | Kuthai Lake | Little Trapper Lake | $9 / 21 / 1990$ | 95 |
|  | 297 | Taku River Mainstem | Kuthai Lake | 2006 | 95 |
|  | 298 | Snettisham Hatchery | Taku River Mainstem | $9 / 24 / 2007$ | 95 |
|  |  | Speel Lake | $9 / 17 / 2003$ | 95 |  |
|  | 299 | Crescent Lake | Snettisham Hatchery | $11 / 27 / 2006$ | 95 |
|  | 300 | Kook Lake | Crescent Lake | $9 / 10 / 2003$ | 94 |
| 301 | Sitkoh Lake | Kook Lake | $7 / 30 / 2007$ | 95 |  |
|  | 302 | Kanalku Lake | Sitkoh Lake | $9 / 26 / 2003$ | 95 |
| 303 | Falls Lake | Kanalku Lake | $7 / 7 / 2007$ | 95 |  |
| 304 | Salmon Lake | Falls Lake | $9 / 2 / 2003$ | 95 |  |
| 305 | Redfish Lake Beaches | Salmon Lake | $7 / 21 / 2007$ | 91 |  |
| 306 | Kutlaku Lake | Redfish Lake Beaches | $8 / 10 / 1993$ | 95 |  |
| 307 | Petersburg Lake | Kutlaku Lake | $9 / 17 / 2003$ | 95 |  |
|  | 308 | Kah Sheets Lake | Petersburg Lake | $8 / 23 / 2004$ | 95 |
| 309 | Tahltan Lake | Kah Sheets Lake | $8 / 25 / 2003$ | 96 |  |
| 310 | Little Tahltan Lake | Tahltan Lake | 9006 | 95 |  |
| 311 | Stikine Devil's Elbow* | Little Tahltan Lake | $9 / 24 / 1990$ | 95 |  |
| 312 | Scud River | Stikine Devil's Elbow | $9 / 7 / 2007$ | 55 |  |
| 313 | Porcupine River* | Scud River | $9 / 13 / 2007$ | 88 |  |
| 314 | Stikine Andy Smith Slough* | Porcupine River | $9 / 13 / 2007$ | 36 |  |
| 315 | Stikine Fowler Slough* | Stikine Andy Smith Slough | $9 / 15 / 2007$ | 10 |  |
| 316 | Craig River* | Stikine Fowler Slough | $9 / 15 / 2007$ | 11 |  |
|  |  | Craig River | 2006 | 12 |  |
|  |  | Craigson Slough | $9 / 14 / 2007$ | 43 |  |

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| Reporting group | Pop \# Population |  | H-W Collection | Date | N |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 335 | Three Mile Creek | Three Mile Creek | 9/30/2004 | 95 |
|  | 336 | Hetta Lake | Hetta Lake | 10/1/2003 | 94 |
|  | 337 | Klakas Lake | Klakas Lake | 9/12/2004 | 95 |
|  | 338 | Kegan Lake | Kegan Lake | 9/10/2004 | 95 |
|  | 339 | Karta River | Karta River | 8/25/1992 | 93 |
|  |  |  | McGilvery Creek | 9/4/2003 | 96 |
|  | 340 | Luck Lake | Luck Lake | 9/10/2004 | 94 |
|  | 341 | Sweetwater Lake | Sweetwater Lake | 6/7/2003 | 47 |
|  |  |  | Sweetwater Lake | 6/23/2007 | 95 |
|  | 342 | Essowah Lake | Essowah Lake | 9/5/2004 | 96 |
|  | 343 | Bowser Lake | Bowser Lake | 9/13/2001 | 95 |
|  | 344 | Damdochax Creek | Damdochax Creek | 9/18/2001 | 94 |
|  | 345 | Tintina Creek | Tintina Creek | 9/12/2006 | 94 |
|  | 346 | Meziadin Lake | Meziadin Lake | 9/19/2001 | 91 |
|  |  |  | Meziadin Beach | 9/26/2006 | 95 |
|  | 347 | Hanna Creek | Hanna Creek | 9/3/2006 | 93 |
|  | 348 | Kitlope Lake | Kitlope Lake | 8/3/2006 | 95 |
|  | 349 | Four Mile Creek | Four Mile Creek | 8/29/2006 | 85 |
|  | 350 | Pinkut Creek | Pinkut Creek | 8/25/2006 | 95 |
|  | 351 | Pierre Creek | Pierre Creek | 8/30/2006 | 95 |
|  | 352 | Fulton River | Fulton River | 2006 | 95 |
|  | 353 | Morrison Arm | Morrison Arm | 9/7/2007 | 92 |
|  | 354 | Lower Tahlo River | Lower Tahlo River | 1988 | 10 |
|  |  |  | Lower Tahlo River | 1994 | 85 |
|  | 355 | Upper Babine River | Upper Babine River | 2006 | 95 |
|  | 356 | Sustut River | Sustut River | 2006 | 95 |
|  | 357 | Slamgeesh River | Slamgeesh River | 8/7/2006 | 95 |
|  | 358 | Swan Lake | Swan Lake | 10/15/2006 | 94 |

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| Reporting group | Pop \# Population | H-W Collection | Date | N |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 359 | Nangeese River* | Nangeese River | $9 / 19 / 2006$ | 42 |  |
|  | 360 | Zymoetz River* | Zymoetz River | $9 / 3 / 2006$ | 64 |
|  | 361 | Nanika River | Nanika River | $9 / 21 / 2007$ | 94 |
|  | 362 | Kitsumkalum Lake* | Kitsumkalum Lake | $11 / 6 / 2006$ | 56 |
|  | 363 | Lakelse Lake | Lakelse Lake | $8 / 22 / 2006$ | 93 |
|  | 364 | Alastair Lake | Alastair Lake | $9 / 14 / 2006$ | 85 |
|  | 365 | Naden River | Naden River | 1995 | 95 |
|  | 366 | Stellako River | Stellako River | $9 / 28 / 2007$ | 94 |
|  | 367 | Horsefly River | Upper Horsefly River | $9 / 2 / 2001$ | 95 |
|  |  | Lower Horsefly River | $9 / 12 / 2001$ | 95 |  |
|  | 368 | Chilko Lake | Chilko Lake | $1 / 1 / 2001$ | 95 |
|  | 369 | Raft River | Raft River | $9 / 4 / 2001$ | 95 |
|  | 370 | Adams River | Adams River | $10 / 3 / 2007$ | 95 |
| 371 | Birkenhead River | Birkenhead River | $10 / 18 / 2007$ | 95 |  |
| 372 | Weaver Creek | Weaver Creek | $1 / 1 / 2001$ | 94 |  |
| 373 | Harrison River | Harrison River | $10 / 17 / 2007$ | 95 |  |
| 374 | Baker Lake | Baker Lake | $5 / 16 / 1996$ | 97 |  |
| 375 | Cedar River | Cedar River | $10 / 26 / 1994$ | 96 |  |
|  |  |  |  | 9,648 |  |

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Table 2. Forty-five sockeye SNP markers assayed for this project; three mitochondrial DNA and 42 nuclear DNA. Forward and reverse primers and probes are given for each new Taqman assay. Loci that were out of $\mathrm{H}-\mathrm{W}$ equilibrium at more than the number of populations expected by chance ( 19 populations @ $P=0.05$ ) are noted with the number of populations out of H-W equilibrium ( $P=0.05$ ) under the H-W column.

| Marker | Reference ${ }^{1}$ | H-W |
| :---: | :---: | :---: |
| One_ACBP-79 | A |  |
| One_ALDOB-135 | A |  |
| One_ctgf-301 | A |  |
| One_CO1 ${ }^{2}$ | A |  |
| One_Cytb_17 ${ }^{2}$ | A |  |
| One_Cytb_26 ${ }^{2}$ | A |  |
| One_E2-65 | B |  |
| One_GHII-2165 | A | 21 |
| One_GPDH-201 | B | 20 |
| One_GPDH2-187 | B |  |
| One_GPH-414 | A |  |
| One_hsc71-220 | A |  |
| One_HGFA-49 | B | 21 |
| One_HpaI-71 | A |  |
| One_HpaI-99 | A |  |
| One_IL8r-362 |  |  |
| F: TTGCTAGAAGCGTTGGTTATGATGA |  |  |
| R: CAGCAAAATTGAGAAGTCACTAGGAAAA |  |  |
| VIC- CAGCCAAAGAAGAGTC |  |  |
| FAM- AGCCAAAAAAGAGTC |  |  |
| One_KPNA-422 | A |  |
| One_LEI-87 | A |  |
| One_MARCKS-241 |  |  |
| F: CCTATCACAGCTTGGTTGAGTTCAA |  |  |
| R: TCCACCCGCTCATTTTTGTAAGAT |  |  |
| VIC-TTGCTTAAAAGGTCTTCC |  |  |
| FAM-TTGCTTAAAAGGTCATCC |  |  |
| One_MHC2_190 ${ }^{3}$ | A | 29 |
| One_MHC2_251 ${ }^{3}$ | A | 30 |
| One_Ots213-181 | A |  |
| One_p53-534 | A |  |
| One_ins-107 | B | 23 |
| One_Prl2 | A |  |

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| Marker | Reference ${ }^{1}$ | H-W |
| :---: | :---: | :---: |
| One_RAG1-103 | A |  |
| One_RAG3-93 | A |  |
| One_RFC2-102 | B |  |
| One_RFC2-285 | B |  |
| One_RH2op-395 | A |  |
| One_serpin-75 | B |  |
| One_STC-410 | A | 22 |
| One_STR07 | A |  |
| One_Tf_ex11-750 | A |  |
| One_Tf_in3-182 | A |  |
| One_U301_92 | A |  |
| One_U401-224 |  | 20 |
| F: GGGTGGAGACGAACGGATTC |  |  |
| R: GTACGATTTTTTTGTAGCCCCAAGT |  |  |
| VIC-CACCTGGAAAGGACTGA |  |  |
| FAM-ACACCTGGAAATGACTGA |  |  |
| One_U404-229 |  |  |
| F: GTTTGTGTGTTGGTGTTTGTCCTT |  |  |
| R: CATTTATCTTGGTGGACGTGTGAGT |  |  |
| VIC-CATGTTCTTCAGTGAACC |  |  |
| FAM-ATGTTCTTCAATGAACC |  |  |
| One_U502-167 |  |  |
| F: GCTTTTGTGCAATAGCTATGTTGCT |  |  |
| R: GCAAAGGTAGGCAGCAGATTG |  |  |
| VIC-CTTCTTGATCAATAACG |  |  |
| FAM-CTTCTTGATCGATAACG |  |  |
| One_U503-170 |  | 20 |
| F: GATTCAGAATTGCCACGACAAAGAA |  |  |
| R: GTGATTGGTACATGTCTGTCGAGTT |  |  |
| VIC-AAGTACTAAAATCAGTTTTACATTG |  |  |
| FAM-TACTAAAATCAGTTGTACATTG |  |  |
| One_U504-141 |  |  |
| F: GCTATAGCTCACAGAGGATCCCA |  |  |
| R: TATTGGCGGGTGAGGGATG |  |  |
| VIC-TCAAGGACACAAACAA |  |  |
| FAM-TCAAGGACAAAAACAA |  |  |
| One_U508-533 |  |  |

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| Marker | Reference ${ }^{1}$ | H-W |
| :---: | :---: | :---: |
| R: CTCAAAGGGTCTGAATACTTATGTAAATAAGGT |  |  |
| VIC-ACACTACAGCCTTATTC |  |  |
| FAM-ACACTACAGCTTTATTC |  |  |
| One_VIM-569 | A |  |
| One_ZNF-61 |  |  |
| F: CCATTCATGTTCTATTCAGATATATTTTGTGCA |  |  |
| R: CCTAGCTAGAGCTCAACAATATGCA |  |  |
| VIC-CTATGGACATGATCTTT |  |  |
| FAM-TTCTATGGACATTATCTTT |  |  |
| One_Zp3b-49 | B |  |

${ }^{1}$ A) Elfstrom et al. (2006); B) Smith et al. (2005).
${ }^{2}$ mtDNA markers; composite haplotype loci were assembled for MSA analyses.
${ }^{3}$ MHC markers were significantly linked in more than $50 \%$ of collections. Composite phenotypes were assembled for MSA analyses.

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Table 3. Descriptive statistics for SNPs used in the current ADF\&G sockeye salmon baseline, including expected $\left(\mathrm{H}_{\mathrm{e}}\right)$ and observed heterozygosity $\left(\mathrm{H}_{\mathrm{o}}\right)$ for nuclear loci, and $\mathrm{F}_{\text {ST }}$ for all nuclear and mitochondrial markers and for the combined nuclear marker. Minimum and maximum values and overall $\mathrm{F}_{\mathrm{ST}}$ are shown, while average heterozygosities include only nuclear markers. Superscripts indicate sets of markers which were pooled into a single locus.

| SNP | $\mathrm{H}_{\mathrm{e}}$ | $\mathrm{H}_{0}$ | $\mathrm{F}_{\text {ST }}$ |
| :---: | :---: | :---: | :---: |
| One_ACBP-79 | 0.472 | 0.406 | 0.121 |
| One_ALDOB-135 | 0.286 | 0.252 | 0.116 |
| One_ctgf-301 | 0.045 | 0.042 | 0.048 |
| One_E2-65 | 0.338 | 0.302 | 0.110 |
| One_GHII-2165 | 0.307 | 0.220 | 0.275 |
| One_GPDH-201 | 0.492 | 0.447 | 0.083 |
| One_GPDH2-187 | 0.210 | 0.172 | 0.168 |
| One_GPH-414 | 0.447 | 0.383 | 0.138 |
| One_hcs71-220 | 0.333 | 0.298 | 0.108 |
| One_HGFA-49 | 0.307 | 0.277 | 0.088 |
| One_HpaI-71 | 0.465 | 0.400 | 0.133 |
| One_HpaI-99 | 0.204 | 0.157 | 0.218 |
| One_ILSr-362 | 0.123 | 0.114 | 0.092 |
| One_KPNA-422 | 0.378 | 0.339 | 0.098 |
| One_LEI-87 | 0.478 | 0.420 | 0.114 |
| One_MARCKS-241 | 0.032 | 0.029 | 0.073 |
| One_MHC2_190 ${ }^{\text {a }}$ | 0.491 | 0.305 | 0.356 |
| One_MHC2_251 ${ }^{\text {a }}$ | 0.491 | 0.334 | 0.303 |
| One_Ots213-181 | 0.277 | 0.241 | 0.125 |
| One_p53-534 | 0.071 | 0.061 | 0.125 |
| One_ins-107 | 0.496 | 0.434 | 0.114 |
| One_Prl2 | 0.500 | 0.447 | 0.096 |
| One_RAG1-103 | 0.055 | 0.050 | 0.102 |
| One_RAG3-93 | 0.160 | 0.143 | 0.104 |
| One_RFC2-102 | 0.348 | 0.307 | 0.112 |
| One_RFC2-285 | 0.099 | 0.088 | 0.100 |
| One_RH2op-395 | 0.018 | 0.017 | 0.042 |
| One_serpin-75 | 0.072 | 0.066 | 0.064 |
| One_STC-410 | 0.456 | 0.353 | 0.220 |
| One_STR07 | 0.460 | 0.393 | 0.145 |
| One_Tf_ex11-750 | 0.488 | 0.387 | 0.206 |
| One_Tf_in3-182 | 0.154 | 0.112 | 0.268 |
| One_U301-92 | 0.277 | 0.252 | 0.089 |
| One_U401-224 | 0.488 | 0.439 | 0.107 |
| One_U404-229 | 0.123 | 0.103 | 0.162 |

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| SNP | $\mathrm{H}_{\mathrm{e}}$ | $\mathrm{H}_{\mathrm{o}}$ | $\mathrm{F}_{\text {ST }}$ |
| :--- | :---: | :---: | :---: |
| One_U502-167 | 0.046 | 0.044 | 0.049 |
| One_U503-170 | 0.254 | 0.224 | 0.115 |
| One_U504-141 | 0.389 | 0.351 | 0.089 |
| One_U508-533 | 0.092 | 0.079 | 0.125 |
| One_VIM-569 | 0.219 | 0.197 | 0.094 |
| One_ZNF-61 | 0.415 | 0.352 | 0.152 |
| One_zP3b-49 | 0.235 | 0.174 | 0.266 |
| One_CO1 ${ }^{\text {b }}$ | $\mathrm{N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 0.254 |
| One_Cytb_17 |  |  |  |
| One_Cytb_26 |  | $\mathrm{N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ |
| One_CO1_Cytb17_26 | $\mathrm{N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 0.498 |
| One_MHC2_190_251 | $\mathrm{N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 0.255 |
| Minimum | $\mathrm{N} / \mathrm{A}$ | $\mathrm{N} / \mathrm{A}$ | 0.295 |
| Maximum | 0.018 | 0.017 | 0.259 |
| Average/Overall | 0.500 | 0.447 | 0.295 |
|  | 0.288 | 0.243 | 0.149 |

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Table 4. Percent of total collections exhibiting significant linkage disequilibrium for the pairs of loci for which disequilibrium was most commonly observed.

|  |  | Significant linkage disequilibrium |  |  |
| :--- | :--- | :--- | :---: | :---: |
| Criteria | Marker pair |  |  | Percentage of <br> total |
| $\mathrm{P}<0.01$ | One_MHC2_190 | One_MHC2_251 | 320 | $55 \%$ |
|  | One_GPDH | One_GPDH2 | 197 | $34 \%$ |
|  | One_Tf_exl0-750 | One_Tf_ex3-182 | 108 | $19 \%$ |
|  | One_RF-112 | One_RF-295 | 43 | $7 \%$ |

Table 5. Log-likelihood $G$ and associated test statistics for the homogeneity of allele frequency log-likelihood ratio tests over all loci across populations within regions and broad regional groupings. Because the number of populations is heterogeneous across regions, we also tabulate $G$ divided by degrees of freedom (df) for each regional level.

| Broad Regions | Regions | $G$ | df | $P$ | \# of pops | $G / \mathrm{df}$ |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: |
| Western Kamchatka | Western Kamchatka | 2,927 | 328 | 0.00 | 9 | 8.92 |
| Eastern Kamchatka | Eastern Kamchatka | 6,376 | 533 | 0.00 | 14 | 11.96 |
| Norton Sound | Norton Sound | 1,417 | 82 | 0.00 | 3 | 17.27 |
|  | Yukon Kuskokwim | 7,685 | 410 | 0.00 | 11 | 18.74 |
|  | Togiak | 1,436 | 164 | 0.00 | 5 | 8.75 |
| Western Bristol Bay | Igushik | 271 | 123 | 0.00 | 4 | 2.21 |
|  | Wood | 3,207 | 738 | 0.00 | 19 | 4.35 |
|  | Nushagak | 3,566 | 328 | 0.00 | 9 | 10.87 |
|  | Western Bristol Bay Total | 16,165 | 1,763 | 0.00 | 48 | 9.17 |
|  | Kvichak | 15,155 | 697 | 0.00 | 18 | 21.74 |
|  | Alagnak | 1,730 | 123 | 0.00 | 4 | 14.07 |
|  | Naknek | 2,954 | 492 | 0.00 | 13 | 6.00 |
| Eastern Bristol Bay | 1,093 | 123 | 0.00 | 4 | 8.89 |  |
|  | Egegik | 608 | 164 | 0.00 | 5 | 3.71 |
|  | Ugashik | 21,540 | 1,599 | 0.00 | 44 | 13.47 |
|  | Eastern Bristol Bay Total | 11,994 | 861 | 0.00 | 22 | 13.93 |
|  | North Peninsula | 11,105 | 779 | 0.00 | 20 | 14.25 |
| Alaska Peninsula | South Peninsula | 23,098 | 1,640 | 0.00 | 42 | 14.08 |
|  | Alaska Peninsula Total | 177,933 | 4,715 | 0.00 | 116 | 37.74 |
| Western Gulf | Western Gulf | 105,112 | 4,018 | 0.00 | 99 | 26.16 |
| Eastern Gulf | Eastern Gulf | 62,220 | 5,084 | 0.00 | 137 | 12.24 |
| WAAP | 354,568 | 14,678 | 0.00 | 375 | 24.16 |  |
| Coastwide Total |  |  |  |  |  |  |

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|  | $90 \%$ Confidence Interval |  |  |
| :--- | :---: | :---: | :---: |
| Region | Estimate | Lower | Upper |
| Western Kamchatka | 0.969 | 0.949 | 0.986 |
| Eastern Kamchatka | 0.956 | 0.933 | 0.978 |
| Norton Sound | 0.946 | 0.913 | 0.973 |
| Yukon Kuskokwim | 0.908 | 0.862 | 0.949 |
| Togiak | 0.946 | 0.898 | 0.980 |
| Igushik | 0.860 | 0.779 | 0.929 |
| Wood | 0.938 | 0.881 | 0.981 |
| Nushagak | 0.912 | 0.862 | 0.954 |
| Kvichak | 0.950 | 0.924 | 0.973 |
| Alagnak | 0.977 | 0.961 | 0.990 |
| Naknek | 0.947 | 0.916 | 0.974 |
| Egegik | 0.913 | 0.864 | 0.954 |
| Ugashik | 0.855 | 0.784 | 0.914 |
| North Peninsula | 0.893 | 0.851 | 0.932 |
| South Peninsula | 0.917 | 0.882 | 0.948 |
| Western Gulf of Alaska | 0.927 | 0.896 | 0.955 |
| Eastern Gulf of Alaska | 0.967 | 0.946 | 0.985 |

Table 6. Proportion of estimates correctly allocated back to reporting group of origin and $90 \%$ confidence intervals for mixtures of 400 fish simulated from baseline populations that contribute to each reporting region ( $100 \%$ simulations) using the program SPAM.

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|  |  | $90 \%$ Confidence Interval |  |
| :--- | :---: | :---: | :---: |
| Region | Estimate | Lower | Upper |
| Western Kamchatka | 0.990 | 0.972 | 1.000 |
| Eastern Kamchatka | 0.974 | 0.934 | 0.996 |
| Norton Sound | 0.985 | 0.961 | 0.999 |
| Yukon Kuskokwim | 0.978 | 0.926 | 0.999 |
| Togiak | 0.987 | 0.960 | 1.000 |
| Igushik | 0.974 | 0.899 | 0.999 |
| Wood | 0.957 | 0.823 | 0.999 |
| Nushagak | 0.956 | 0.866 | 0.998 |
| Kvichak | 0.959 | 0.901 | 0.998 |
| Alagnak | 0.992 | 0.973 | 1.000 |
| Naknek | 0.972 | 0.933 | 0.997 |
| Egegik | 0.947 | 0.868 | 0.995 |
| Ugashik | 0.959 | 0.898 | 0.996 |
| North Peninsula | 0.980 | 0.935 | 0.999 |
| South Peninsula | 0.958 | 0.914 | 0.991 |
| Western Gulf of Alaska | 0.894 | 0.827 | 0.948 |
| Eastern Gulf of Alaska | 0.983 | 0.950 | 0.999 |

Table 7. Proportion of estimates correctly allocated back to reporting group of origin and $90 \%$ credibility intervals for mixtures of 200 known fish that were removed from the baseline populations that contribute to each reporting region ( $100 \%$ proof tests) using the program BAYES with a flat prior.

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Figure 1. Locations where sockeye salmon were sampled for tissues suitable for genetic analysis from throughout the Pacific Rim. These tissues were screened for 42 nuclear and 3 mitochondrial single nucleotide polymorphism markers. This baseline, augmented with additional markers, will serve as a baseline to examine the potential power and precision of stock composition estimates from fishery samples taken under the Western Alaska Salmon Identification Program. Colors denote eight geographic regions that match the colors and regions in Figure 6. Western and Eastern Kamchatka, Norton Sound, and Eastern and Western Gulf of Alaska represent five of the proposed reporting groups. The remaining regions (Western Bristol Bay YK, Eastern Bristol Bay, and the Alaska Peninsula) are further subdivided into a total of 12 reporting groups as shown in Figures 2 and 7.


Figure 2. Sockeye salmon sample locations from Western Alaska and the Alaska Peninsula (WAAP) included in the SNP baseline. Colors denote the 13 WAAP reporting regions.


Figure 3. Number of loci that were out of H-W equilibrium $(P=0.05)$ for 0 to 30 populations. By chance, the one would expect 18.75 populations to be out of $\mathrm{H}-\mathrm{W}$ expectation at this criterion (375 populations * 0.05 ). We review the loci that were out of $\mathrm{H}-\mathrm{W}$ equilibrium at more that 23 populations in the text.

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Figure 4. Number of baseline populations that were out of $\mathrm{H}-\mathrm{W}$ equilibrium $(P=0.05)$ for 0 to 12 loci. By chance, the one would expect 2.1 loci to be out of $\mathrm{H}-\mathrm{W}$ expectation at this criterion (i.e., 42 loci $* 0.05$ ). We review the populations that were out of $\mathrm{H}-\mathrm{W}$ equilibrium at more that 5 loci in the text.

Fst/He


Figure 5. LOSITAN (Antao et al. 2008) graphical output showing the relationship between $\mathrm{F}_{\text {ST }}$ and $\mathrm{H}_{\mathrm{e}}$ for SNP markers analyzed in select populations from western Alaska and the north Alaska Peninsula (method details in text). The expected distribution of $\mathrm{F}_{S T}$ and $\mathrm{H}_{\mathrm{e}}$ under an island model of migration with neutral markers is shown in gray. Loci in the red area are candidates for positive selection and loci in the yellow area are candidates for balancing selection. Outlier loci are tagged with labels.

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Figure 6. Unweighted pair-group method (UPGMA) tree of Cavalli-Sforza and Edwards chord distances among the 375 populations included in the coastwide 42 SNP baseline. Population numbers correspond to those in Table 1. Note the high variation within the Gulf of Alaska relative to the WAAP.

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Figure 7. Unweighted pair-group method (UPGMA) tree of Cavalli-Sforza and Edwards chord distances among the 137 populations included in the WAAP portion of the coastwide 42 SNP baseline.


Figure 8. Log-likelihood ratio test statistics $(G)$ divided by degrees of freedom (df) over all loci by reporting group within the WAAP area.


Figure 9. Log-likelihood ratio test statistics $(G)$ divided by degrees of freedom (df) over all loci by region within the full baseline.


Figure 10. Log-likelihood ratio test ( $G$ ) statistics divided by degrees of freedom (df) for each SNP marker for the populations within the full coastwide baseline and the more restricted WAAP baseline. Note the similar and high values for the $G$ statistics for both geographic regions at the one MHC marker included in this analysis and the generally lower values for the $G$ statistics in the WAAP area for the remaining markers.


Figure 11. Proportion of estimates correctly allocated back to reporting group of origin and $90 \%$ confidence intervals for mixtures of 400 fish simulated from baseline populations that contribute to each reporting region ( $100 \%$ simulations) using the program SPAM.


Figure 12. Proportion of estimates correctly allocated back to reporting group of origin and $90 \%$ credibility intervals for mixtures of 200 known fish that were removed from the baseline populations that contribute to each reporting region ( $100 \%$ proof tests) using the program BAYES with a flat prior.


[^0]:    ${ }^{1}$ This document serves as a record of communication between the Alaska Department of Fish and Game Commercial Fisheries Division and the Western Alaska Salmon Stock Identification Program Technical Committee. As such, these documents serve diverse ad hoc information purposes and may contain basic, uninterpreted data. The contents of this document have not been subjected to review and should not be cited or distributed without the permission of the authors or the Commercial Fisheries Division.

